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An Examination of the Relationship between Levels of Drinking Water Quality and the Occurrence of Self-Reported Diarrheal Disease: A Prospective Cohort Study in the Dominican Republic, 2005-2006.

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Abstract

Shannon M. Kraft

An Examination of the Relationship between Levels of Drinking Water Quality and the Occurrence of Self-Reported Diarrheal Disease: A Prospective Cohort Study in the Dominican Republic, 2005-2006.

(Under the direction of Christine Stauber, Faculty Member)

Background: 884 million people do not have access to clean water, which is a potential contributor to diarrhea (JMP, 2010). The purpose of this study was to examine the potential associations between the occurrence of diarrhea and the levels of turbidity, total coliforms, and *Escherichia coli* (*E. coli*), in 185 households in Bonao, Dominican Republic in 2005-2006.

Methods: A biweekly water quality dataset and a weekly diarrhea occurrence dataset were merged using three different methods. T-tests and odds ratios were calculated for all three different datasets. Multivariate logistic regression was also conducted.

Results: There were 430 cases of diarrhea out of 14,245 observations. In the age-adjusted multivariate logistic regression, turbidity (OR = 1.36; p-value = .012) and total coliforms (OR = .842; p-value = .006) were found to be significant for an association with the occurrence of diarrhea. *E. coli* was not found to be significant for an association.

Conclusions: This study strengthens the evidence supporting a positive association between turbidity and the occurrence of diarrhea. This study also showed a negative association between total coliforms and diarrhea. Future studies are needed to clarify these associations.

INDEX WORDS: diarrhea, water quality, turbidity, total coliforms, *E. coli*, Dominican Republic

AN EXAMINATION OF THE RELATIONSHIP BETWEEN LEVELS
OF DRINKING WATER QUALITY INDICATORS AND THE
OCCURRENCE OF SELF-REPORTED DIARRHEAL DISEASE: A
PROSPECTIVE COHORT STUDY IN THE DOMINICAN
REPUBLIC, 2005-2006.

By SHANNON M. KRAFT

B.S., BERRY COLLEGE

A Thesis Submitted to the Graduate Faculty of Georgia State University in Partial Fulfillment of
the Requirements for the Degree

Master of Public Health

Atlanta, GA 30303

AN EXAMINATION OF THE RELATIONSHIP BETWEEN LEVELS OF DRINKING
WATER QUALITY AND THE OCCURRENCE OF SELF-REPORTED DIARRHEAL
DISEASE: A PROSPECTIVE COHORT STUDY IN THE DOMINICAN REPUBLIC, 2005-
2006.

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CHAPTER I

INTRODUCTION

1.1 Background

Based on the 2010 Joint Monitoring Programme Report, 884 million people lack access to clean water, and 2.6 billion do not have adequate sanitation (JMP, 2010). It is estimated that about 94% of all diarrheal disease is caused by a “reasonably modifiable environment”, which implies that improving this environment through increased access to clean water and sanitation could decrease the occurrence of diarrheal disease (Pruss-Ustun *et al.* 2007). In fact, it has been shown that there is a correlation between improving access to clean water and improved sanitation and a reduction in the occurrence of diarrheal disease (WHO, 2000). Since the 1980s, the number of deaths due to diarrheal disease has fallen from 4.6 million in 1982 to 2.5 million as of 2003. However, morbidity from diarrheal diseases has not decreased (Kosek, 2003).

According to the World Health Organization (WHO), 95% of the population of the Dominican Republic has access to improved water and 78% have access to improved sanitation (WHO, 2009). Despite the improvements in access to water and sanitation in the Dominican Republic, improved access does not always mean that the water is safe to drink.

1.2 Purpose of Study

The purpose of this study was to determine if there are any relationships between increased levels of *E. coli*, total coliforms, or turbidity in household drinking water and the frequency of self-reported diarrheal disease for participants in households in Bonao, Dominican Republic.

1.2 Research Questions

The purpose of this study was to determine if there is an association between water quality and the frequency of self-reported diarrheal disease among individuals in Bonao, Dominican Republic. Potential associations were determined by answering the following questions:

- 1) Is there an association between increased levels of *E. coli* in household drinking water and self-reported frequency of diarrhea?
- 2) Is there an association between increased levels of total coliforms in household drinking water and self-reported frequency of diarrhea?
- 3) Is there an association between increased turbidity in household drinking water and self-reported frequency of diarrhea?

CHAPTER II REVIEW OF THE LITERATURE

The purpose of this study was to look at the global burden of diarrheal disease on developing countries, particularly the Dominican Republic, to analyze the efficacy of current bacterial indicators, and to review previous studies on potential associations between diarrheal disease and water quality.

2.1 The Dominican Republic

The Dominican Republic is located on the western half of the island of Hispaniola, which is located in the Caribbean Sea (PAHO, 2010). It has a land area of 48,670 sq. km. (CIA, 2009). The Dominican Republic borders the country of Haiti (PAHO, 2010). It has a tropical climate with an average annual temperature of 26 °C (78 °F) (CIA, 2009; UN, 2010).

The Dominican Republic has a population of 9,650,054 as of July 2009. The birth rate is 22.39 births/1000 with an infant mortality rate of 25.96/1000 live births. The average life expectancy is 73.7 years at birth (CIA, 2009). According to the World Health Organization (WHO), 20% of the population lives below the poverty line (WHO).

The World Bank categorizes countries based on their gross national income (GNI) per capita. Based on GNI, a country will be classified as low-income, middle-income, and high-income. Low and middle-income countries are arbitrarily classified as developing countries. According to the World Bank, the Dominican Republic was classified as an upper middle-income country with a GNI per capita of \$4,390 in 2008. Ergo, for the purposes of this study, the Dominican Republic can be arbitrarily considered a developing country (World Bank, 2009).

Although the Dominican Republic is classified as a developing country, it is more developed than many other developing nations. For instance, when the Dominican Republic is

compared to its neighbor, Haiti, the differences between the two countries are startling. Haiti occupies the eastern half of Hispaniola. Haiti has a population of 9,446,000 with a live birth rate of 29.1/1000 live births (WHO, 2008; CIA, 2009). The average life expectancy is 73.7. Additionally, 80% of the population lives below the poverty line (CIA, 2009).

2.2 The Global Burden of Diarrheal Disease Due to Lack of Access to Clean Water and Sanitation

Since the end of the 20th century, the burden of diarrheal disease related to poor water and poor access to sanitation has been recognized (Rheingans *et al.*, 2006). Currently, the World Health Organization estimates that around 2.2 million people die annually from gastrointestinal disease including diarrhea (WHO, 2010). The majority of these deaths occur in children under the age of five with 1.5 million children dying from diarrheal disease annually (UNICEF, 2009).

During the end of the 20th century, several different initiatives were started to address this problem. First, the United Nations (UN) made a push towards improving these conditions by naming the 1980s, the International Decade for Clean Drinking Water. The main goal of the International Decade for Clean Drinking Water was to provide safe drinking water universally. Unfortunately, this goal was not met (Rheingans *et al.*, 2006).

Thus, in 2000, the UN developed the Millennium Development Goals (MDGs) with Goal 7 focusing on decreasing by half the number of people without access to clean water and good sanitation by 2015. As further incentive to improve global access to water, 2005-2015 has been designated as the International Decade for Action-Water for Life (Rheingans *et al.*, 2006). Based on the most recent Joint Monitoring Programme report (2010), the world is expected to exceed Goal 7 by 2015 (JMP, 2010).

However despite this achievement, based on the predicted population growth, an estimated 34-76 million people will still die from diarrheal disease between 2000 and 2020 even with the MDG being met (Gleick *et al.*, 2002). One major contributor to the ongoing occurrence of diarrheal disease and death from diarrheal disease is the lack of access to clean water and good sanitation in many countries (Rheingans *et al.*, 2006). In fact, about 94% of all diarrheal disease is caused by a “reasonably modifiable environment”, which implies that improving this environment by increasing access to clean water and sanitation could decrease the occurrence of diarrheal disease (Pruss-Ustun *et al.*, 2007). Clean water access is defined as regular access to at least 20 liters of water from an improved source per person daily (Rheingans *et al.*, 2006).

Studies have been done that have highlighted the importance of access to clean water and good sanitation (Ako Ako *et al.*, 2009; Gundry *et al.*, 2004). One such study was conducted by Ako Ako *et al.* (2009) in Cameroon. In this particular study, the researchers were interested in the water quality and sanitation and types of waterborne disease in Bonaberi, Cameroon.

Ako Ako *et al.* found that none of the collected samples met the WHO recommendations for bacteriological water quality or turbidity. Additionally, it was found that diarrhea was the most common waterborne disease among children under the age of one year old. Amongst this age group, typhoid fever was the most uncommon. Among children ages 1-4 years old, all four waterborne diseases occurred. In individuals older than 15 years old, typhoid fever was the most common disease (Ako Ako *et al.* 2009).

A meta-analysis conducted by Gundry *et al.* (2004) looked at the impact of interventions that looked at the water storage and water treatment practices and their effects on health (Gundry *et al.* 2004). The authors found that point-of-use interventions were more successful in reducing the occurrence of diarrheal disease when adequate sanitation was practiced within the households

(Gundry *et al.*, 2004). A point-of-use intervention is some form of water treatment that is performed on water at the point of use, such as filtration (Luby *et al.*, 2008). Gundry *et al.* speculated that this could be because the use of adequate sanitation practices resulted in a diminishing of fecal matter that could be transmitted (Gundry *et al.*, 2004).

Unfortunately, many developing countries do not have universal access to either clean water or sanitation. Globally, 884 million people lack access to clean water, and 2.6 billion do not have adequate sanitation (JMP, 2010). The majority of these people are located in Sub-Saharan Africa and Southeast Asia. However, in many middle-low-income countries, only around 79% of the population has clean water and 49% have adequate sanitation (Rheingans *et al.*, 2006). In contrast to other developing nations, 95% of the population of the Dominican Republic has access to improved water sources, and 78% have adequate sanitation (WHO, 2009). Even with the MDG being met, 672 million people will still not have access to clean water in 2015 (JMP, 2010).

2.3 The Efficacy of Bacterial Indicators

Since the 19th century, coliforms have been used as indicators of water quality (Rompre *et al.*, 2002). Coliforms are traditionally the bacteria that are found in the intestines of humans and other animals, which are excreted in their feces (Rompre *et al.*, 2002). For instance, one common coliform is *Escherichia coli* (*E. coli*), which makes up about 1% of the total bacterial population within the human large intestine. *E. coli* is also the predominant bacteria (94%) in the intestines of 80 different farm animals and pets (Leclerc *et al.*, 2001).

Originally, the coliform group consisted of only *Klebsiella pneumoniae*, *Klebsiella rhinoscleromatis*, *E. coli*, *Enterobacter aerogenes*, *Enterobacter cloacae*, and *Citrobacter freundii* (Leclerc *et al.*, 2001). However, the current technical definition of a coliform is either “all aerobic and facultative anaerobic, Gram-negative, non-spore-forming, rod-shaped bacteria that ferment lactose with gas and acid formation within 48 hours at 35° C” or “all aerobic and many facultative anaerobic, Gram-negative, non-spore-forming, rod-shaped bacteria that develop a red colony with a metallic sheen within 24 hours at 35° C on an Endo-type medium containing lactose” (Rompere *et al.*, 2002).

The WHO currently classifies coliforms as part of the group of fecal indicators that can infer the potential presence of pathogens (WHO, 2001). WHO uses *E. coli* to only infer the presence of pathogens because some research (Solo-Gabriele *et al.*, 2000; Anderson *et al.*, 2005) has shown that *E. coli* can be present in subtropical and tropical waters and sediments without there being corresponding fecal contamination. Solo-Gabriele *et al.* (2000) found that *E. coli* levels could multiply in dry soil in a tidally-influenced area in a subtropical climate, which they speculated could influence the accuracy of fecal indicators (Solo-Gabriele *et al.*, 2000). The study conducted by Anderson *et al.* (2005) showed that fecal indicators had slower decay rates in freshwater and sediments, which could also impact the use of fecal indicators as measures of water quality (Anderson *et al.*, 2005).

Another measure commonly used to evaluate water quality is turbidity. Turbidity is defined as a measure of the clarity of water caused by the presence or absence of suspended matter in the water. While increased turbidity could be caused by an increase of bacterial growth, such as total coliforms or *E. coli*, it can also be caused by any other particle in the water (EPA, 1999).

Mann *et al.* (2007) conducted a systematic review of several studies (Schwartz *et al.*, 1997; Morris *et al.*, 1998; Schwartz *et al.*, 2000; Aramini *et al.*, 2000; Lim *et al.*, 2002; Gilbert *et al.*, 2006) to determine if there was any association between drinking water quality and turbidity levels. The results among the six different studies varied depending on the study's setting. Two of the studies (Lim *et al.*, 2002; Morris *et al.*, 1998) found no association, while the other four studies (Schwartz *et al.*, 1997; Schwartz *et al.*, 2000; Aramini *et al.*, 2000; Gilbert *et al.*, 2006) did.

2.4 Potential Associations between Water Quality and Diarrheal Disease

In the past two decades, several studies have been undertaken in different developing locales to study the potential association between water quality and the occurrence of diarrheal disease (Moe *et al.* 1991; Gundry *et al.* 2004; Jensen *et al.* 2004; Brown *et al.* 2008). The results of these studies have been mixed. No study has been able to show a very definitive association between water quality and diarrheal disease.

One of the earlier studies was a one-year study in Cebu, Philippines in 1991 that was conducted by Moe *et al.* (1991). This study looked at the efficacy of the four main bacterial indicators of water quality, fecal coliforms, *E. coli*, enterococci, and fecal streptococci. This study only showed a significant difference in the amount of diarrheal disease when the water quality was bad. They speculated that diarrheal disease among children drinking good to moderate water may be the result of other transmission routes. It appeared that only when the water was severely contaminated did water become the major source of transmission of diarrheal disease (Moe *et al.* 1991).

Gundry et. al (2004) conducted a meta-analysis on studies between 1994 and 2001 that looked at the potential relationship between microbiological water quality indicators at point-of-use to health outcomes. It was found that there was not any significant association between the levels of indicator bacteria and diarrheal disease. This lack of association was true regardless of the severity of the diarrheal disease in any of the studies, which contrasted Moe's results (Gundry *et al.* 2004; Moe *et al.* 1991).

Jensen et. al (2004) conducted a one –year study in southern Punjab, Pakistan. Similar to the studies from Gundry's meta-analysis (Gundry *et al.* 2004), Jensen's study did not find any association between the amount of *E. coli* contamination and the incidence of diarrhea among young children, which remained even as the levels of *E. coli* increased. They concluded that their study supported the theory that water contamination was not as significant of a risk factor for childhood diarrhea in comparison to other risk factors, such as access to water (Jensen *et al.* 2004).

Brown et. al conducted a 22-week study in Kandal Province, Cambodia to examine the association between *E. coli* levels in household drinking water and diarrheal disease and dysentery (2008). This study found that there were small, nonlinear, yet significant increases in the occurrence of diarrheal disease in the groups with exposure levels of 11-100 *E. coli* cfu/100 mL, 101-1000 *E. coli* cfu/100 mL and 1001 + *E. coli* cfu/100 mL. These results were found in the entire group and also in a child only group (Brown *et al.* 2008).

Chapter III

METHODOLOGY

3.1 Data Sources

The data used in this study was secondary data obtained from Dr. Christine Stauber. The dataset was created using data from two different files of data from a prospective cohort study that was conducted between September 2005 and January 2006. The protocol was approved by Georgia State University's Institutional Review Board (protocol H10061).

3.2 Study Population

During September 2005 to January 2006, 185 households in Bona, Dominican Republic were selected to participate in a prospective cohort study as part of a randomized controlled trial of the concrete biosand filter, a household water treatment device. For five months prior to the randomized controlled trial study, households were asked to answer questions regarding diarrheal disease as well as provide drinking water samples to provide a baseline on the occurrence of diarrheal disease and the water quality. The baseline data are the subject of the study here.

During this time period, the households were surveyed weekly regarding the occurrence of diarrheal disease in their households. Additionally, water samples from all household drinking sources were collected and tested for *E. coli*, total coliforms, and turbidity levels twice a month.

3.3 Study Measures

The outcome assessed in this study was the occurrence of diarrheal disease at the individual-level among the studied households. The three main exposures examined were the total coliforms, *E. coli*, and turbidity levels of the household drinking water samples. All three variables were log-transformed to obtain more normal distributions.

The dataset was created from two different datasets. The first dataset contained the weekly individual-level diarrheal disease data. This data included information on the nature of the diarrheal episode, including duration, stool type, treatment, and hospitalization. The second dataset consisted of the results of the biweekly water quality analysis, which included information on total coliform, *E. coli*, and turbidity levels of the household drinking water samples. As shown in Figure 1, data were collected at different frequencies and time points.

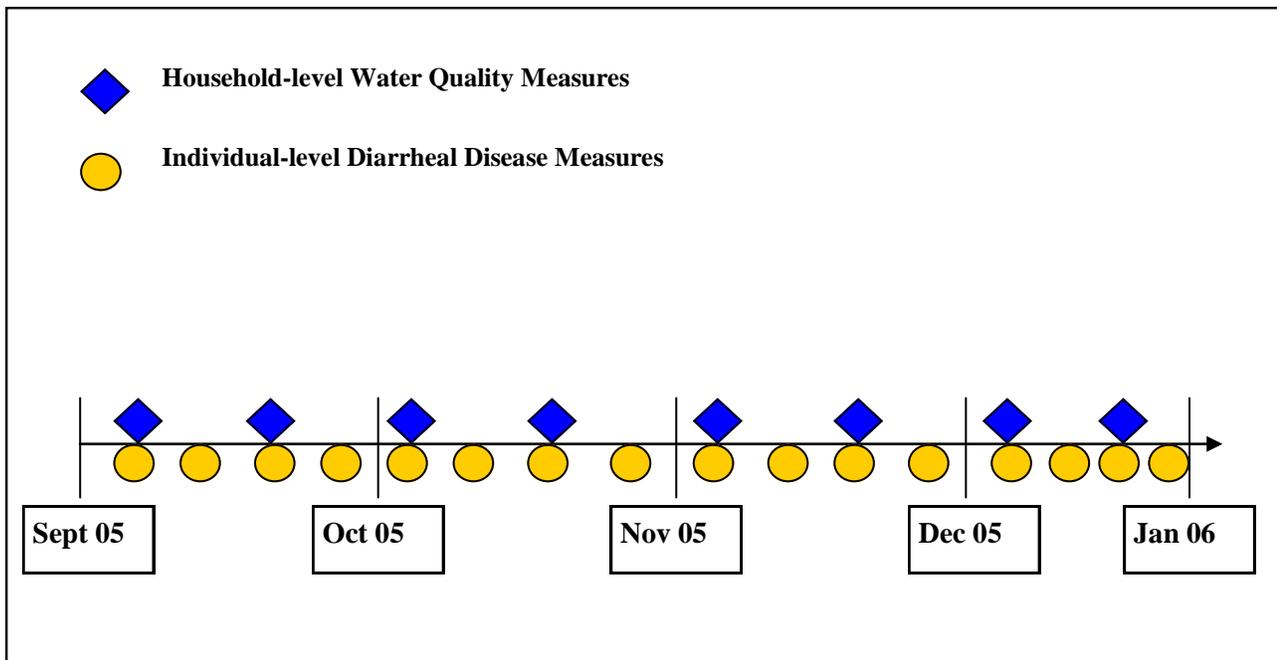


Figure 1: The Collection Times of the Two Datasets

In order to account for the difference between the weekly diarrheal disease measurements and the biweekly water quality measurements, the diarrheal disease data and the water quality data were initially merged together in two ways:

- 1) All diarrheal disease data were directly linked with water quality measurements from the same week of observation. This assumed that any development of diarrheal disease was a direct result of the water quality during the same week.
- 2) A one week lag was incorporated into the water quality data. This resulted in the water quality data being linked with the diarrheal disease data one week after the water quality sample was taken. This accounted for the possible incubation time of diarrheal disease after exposure to potentially contaminated water occurred (Figure 2).

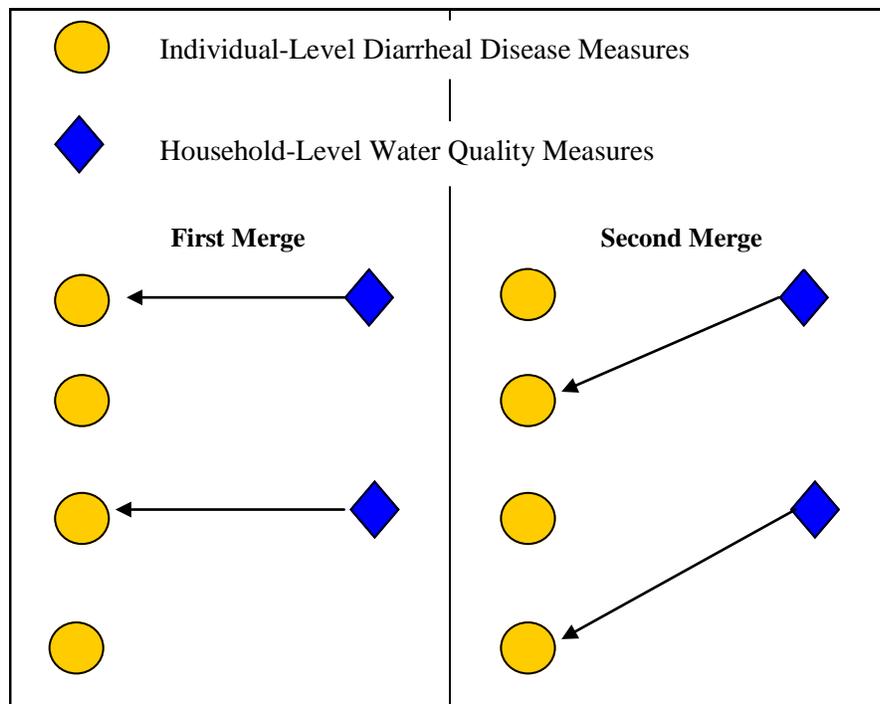


Figure 2: Diagram of the First and Second Data Merges

In either of the above analyses, since water quality was only measured at two week intervals and diarrheal disease was measured at weekly intervals, the merged data were limited to approximately half of the total observations. Therefore, to be able to incorporate water quality and all of the diarrheal disease measurements, a third approach was taken to link the two:

- 3) Water quality data were averaged for each household for each month and then assigned to the corresponding diarrheal disease data for that month. This enabled all of the diarrheal disease observation to be incorporated into further analyses (Figure 3).

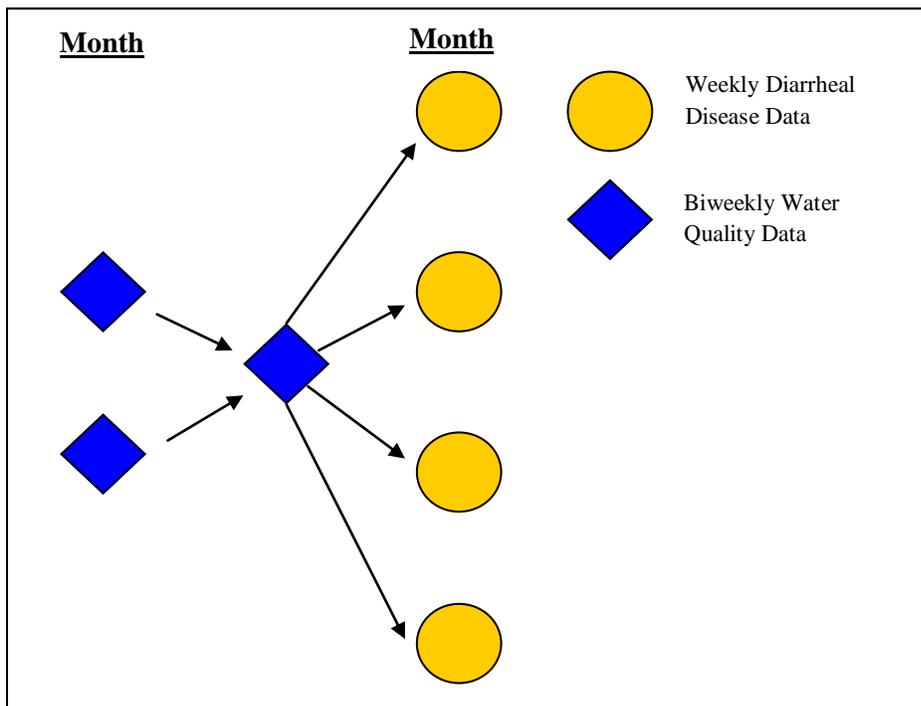


Figure 3: Diagram of Third Data Merge

3.4 Analysis

All data was analyzed in StataSE 10. Descriptive statistics were run on the water quality dataset to determine the designated users of the water sources, the types of water treatment used, turbidity levels, total coliforms levels, *E. coli* levels, occurrence of diarrhea, type of stool, hospitalization, and interference with daily activities.

T-tests were performed on data from all three merges to determine if there were any significant differences between the levels of turbidity, total coliforms, and *E. coli* among people with diarrhea compared to people without diarrhea. Univariate logistic regressions were run to determine the extent of any possible associations between the outcome of diarrheal disease and any of the different measurement variables of total coliforms, *E. coli*, and turbidity. Finally, multivariate logistic regression was performed on the monthly averages dataset.

Chapter IV

RESULTS

The research questions are answered in detail in section IV.

4.1 Water Quality Dataset

The water quality dataset was a biweekly dataset that collected information about a household's water treatment practices. Additionally, information about the turbidity, total coliform levels, and *E. coli* levels were also measured biweekly for each household. There were 185 households that participated, which resulted in 1634 observations.

During the data collection, every household was asked to provide a sample of all sources of household drinking water. It was possible for a household to have more than one water sample associated with it at any given week. Normally, this situation occurred when a household used one water source for adults and older children and another source for the younger children. Additionally, the number of samples associated with a household could vary from week to week.

After providing a water sample from all water sources being used during that particular week, the households were asked to identify whether the source was for adult use, child use, or both. The majority of the water sources were used by both children and adults (N = 1050; 64.26%). Out of the sources, 20.56% of the sources were used by children only (Figure 4).

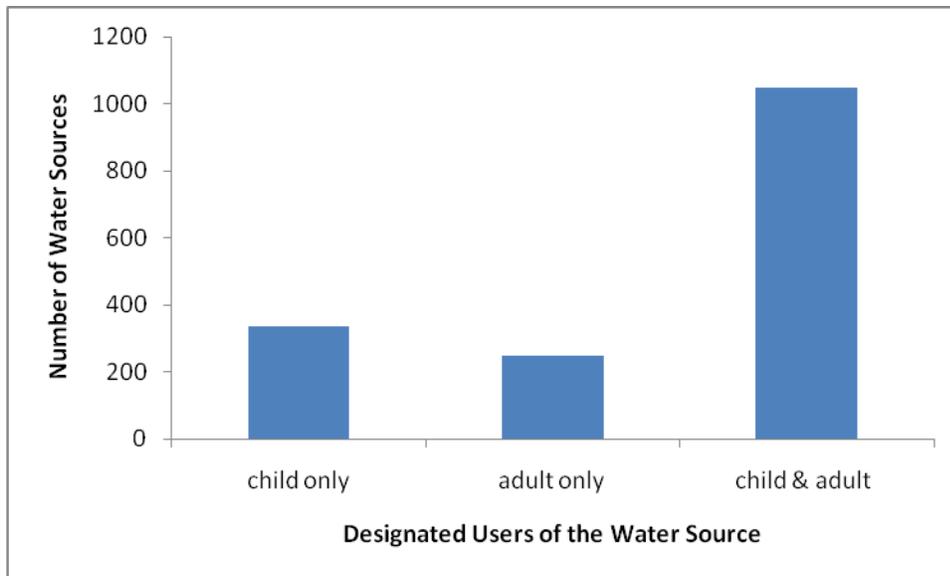


Figure 4: The Number and Type of Water Source

The continuous turbidity levels were \log_{10} -transformed. After the log-transformation, the histogram of turbidity showed a very normal distribution. Around 10% of the observations occurred around $-0.1 \log_{10}$ NTU, which was the highest peak in the histogram (Figure 5).

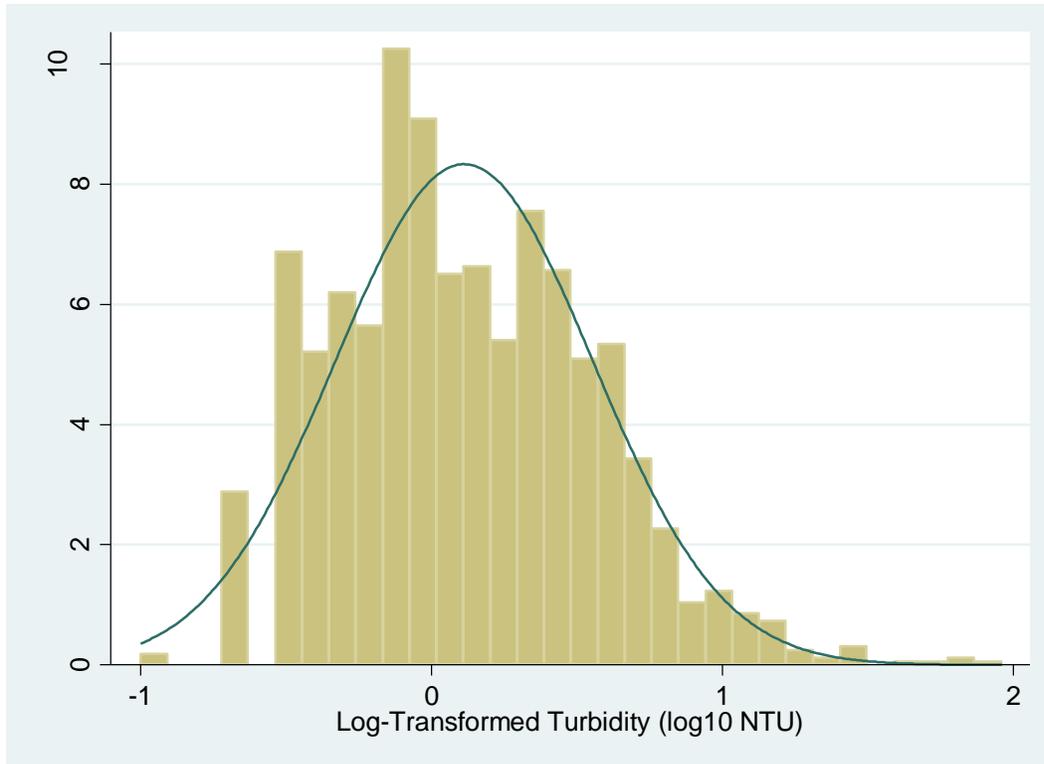


Figure 5: Histogram of the Log-Transformed Turbidity (\log_{10} NTU) Levels

The total coliform levels were also \log_{10} -transformed. After the transformation, the histogram of total coliforms was mainly flat with one very significant peak around $2.5 \log_{10}$ MPN/100 mL. Sixty percent of the observations were clustered around this point (Figure 6).

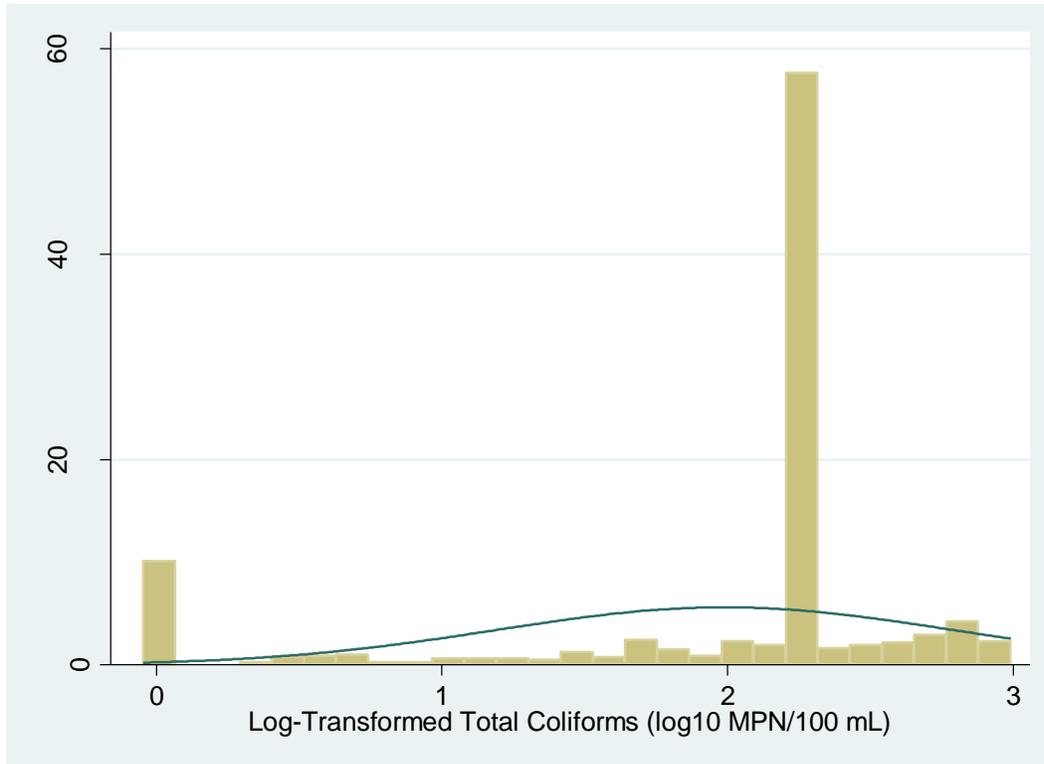


Figure 6: Histogram of Log-Transformed Total Coliforms (\log_{10} MPN/100 mL) Levels

Finally, the *E. coli* levels were also \log_{10} -transformed. After the transformation, the histogram of *E. coli* appeared somewhat normally distributed. However, it has less kurtosis than the histogram of turbidity. The most noteworthy aspect of the *E. coli* histogram was that around 25% of the observations were clustered around 0 or less than 0 \log_{10} MPN/100 mL (Figure 7).

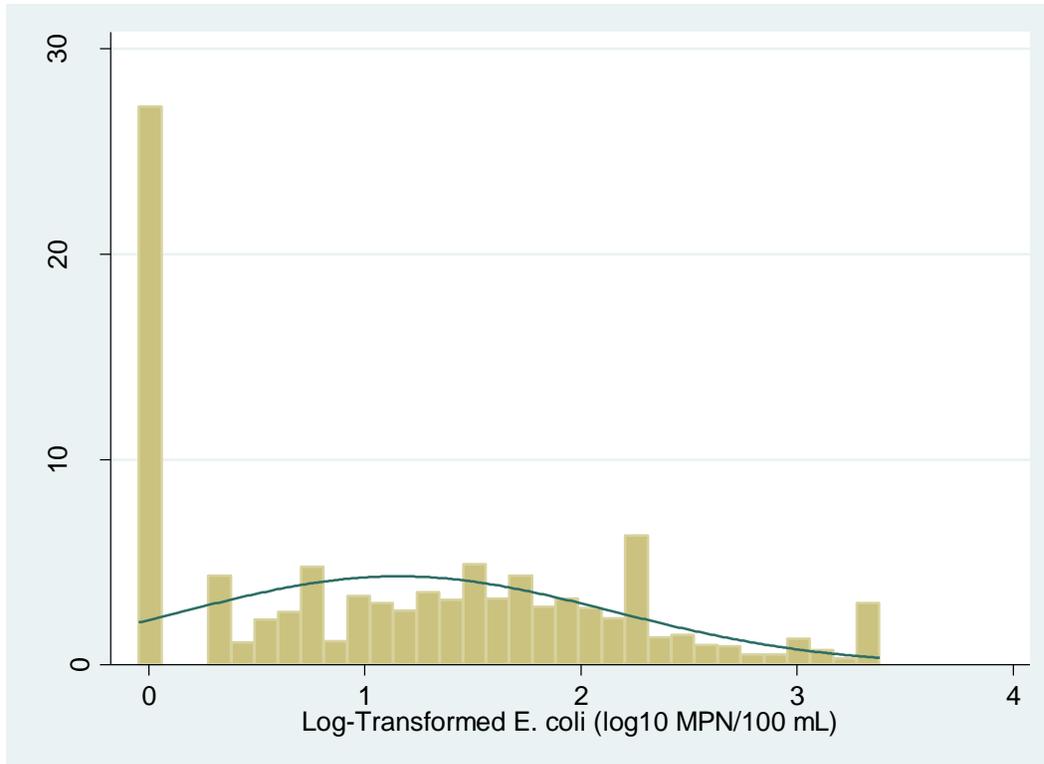


Figure 7: Histogram of Log-Transformed *E. coli* (\log_{10} MPN/100 mL) Levels

After all three exposures were \log_{10} -transformed, box plots of turbidity, total coliforms, and *E. coli* levels during the five-month study period were created (Figure 8; Figure 9).

Turbidity levels were relatively constant during September through November. A slight peak in the levels occurred in December, which was followed by a marked decrease in January. Additionally, there were only a few outliers (Figure 8).

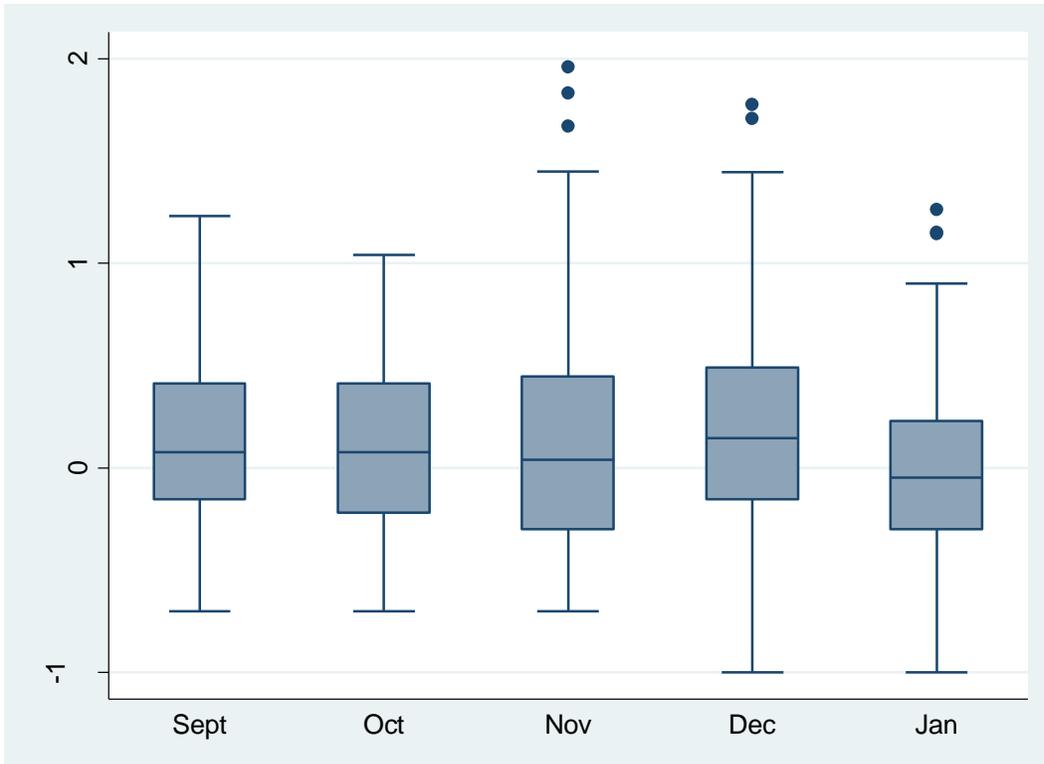


Figure 8: Log-Transformed Turbidity (\log_{10} NTU) Levels during the Study Period

During the study period, *E. coli* levels remained relatively constant with two slight decreases in November and January. There were also no outliers occurring. In contrast, the majority of the total coliform measurements were outliers. However, based on the means, the levels increased in October, then remained constant throughout January (Figure 9).

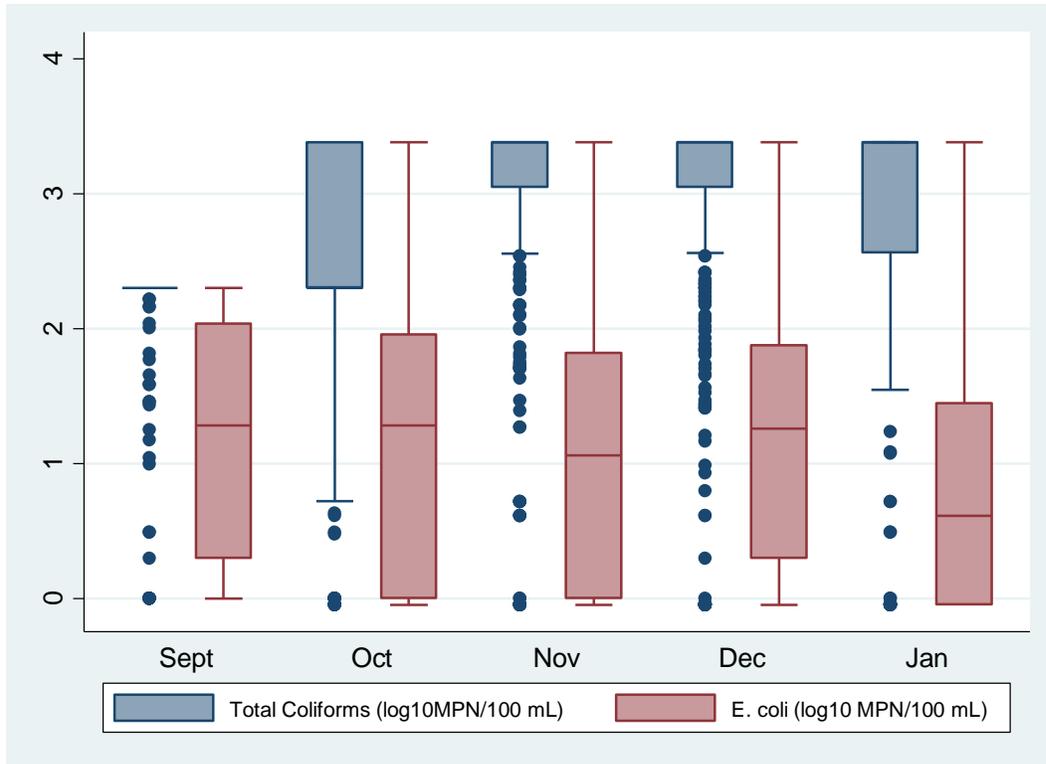


Figure 9: Log-Transformed Total Coliforms (\log_{10} MPN/100 mL) and *E. coli* (\log_{10} MPN/100 mL) Levels during the Study Period

Every household was also asked if they treated their drinking water. The biweekly water quality surveys found that the majority of the provided water samples were not treated (68.8%). However, among the treated water samples, the most commonly used water treatment was boiling (Table 1).

Table 1: Types of Water Treatment Used

Variable (N)	Frequency (%)
Treated (1632)	
<i>Yes</i>	509 (31.2)
<i>No</i>	1123 (68.8)
Boiled (1633)	
<i>Yes</i>	381 (23.3)
<i>No</i>	1252 (76.7)
Bleach (1634)	
<i>Yes</i>	116 (7.10)
<i>No</i>	1518 (92.9)
Other (1634)	
<i>Yes</i>	18 (1.10)
<i>No</i>	1616 (98.9)

Categories are not mutually exclusive

4.2 Diarrheal Disease Dataset

There were 14,297 observations in the diarrheal disease dataset. An observation was any incidence of reported diarrhea for any member of a study household. Thus, a household could have numerous observations associated with it in a study week. In this study, diarrheal disease was classified based on the WHO standard, which defines three or more loose stools or any bloody stools within 24 hours (Stauber *et al.* 2009). Fifty-two of those observations were classified as “I don’t know” in regards to the occurrence of diarrheal disease. Therefore, those 52 observations were not included in any of the analyses performed. Out of the remaining 14,245 observations, there were 430 occurrences of diarrheal disease during the study time period,

which is 3.02% of the observations. This is a relatively low frequency of diarrhea, which could affect the analysis.

There were 290 reported cases of watery diarrhea (69.05%) making it the most common type of diarrhea experienced. The majority of the reported diarrheal disease cases (152) did not interfere with the individual's daily activities (62.83%). Nine cases of diarrhea resulted in hospitalization (2.16%). (Table 2).

Table 2: Descriptive Statistics of Diarrhea among the Study Population

Variable (N)	Frequency (%)
Diarrhea (14,245)	
<i>Yes</i>	430 (3.02)
<i>No</i>	13,815 (96.98)
Type of Stool (420)	
<i>Watery</i>	290 (69.05)
<i>Bloody</i>	12 (2.86)
<i>Mucous</i>	96 (22.86)
<i>Other</i>	22 (5.24)
Interference with Daily Activities (417)	
<i>Yes</i>	152 (36.45)
<i>No</i>	262 (62.83)
<i>I don't know</i>	3 (0.72)
Hospitalized (416)	
<i>Yes</i>	9 (2.16)
<i>No</i>	407 (97.84)

Out of 14,297 observations, the age group with the greatest percentage of diarrheal disease among the same population was the <2 year olds with 176 cases of diarrhea (12.9%). In the next age group, 2-4 years old, the percentage of diarrheal disease decreased by roughly half from the <2 years old group. Overall, as age increased, the frequency of diarrhea decreased as well (Table 3).

Table 3: Descriptive Statistics of Diarrhea by Age Group among the Sample Unit

Age (N)	Diarrhea (%)	No Diarrhea (%)
< 2 years old (1362)	176 (12.9)	1186 (87.1)
2-4 years old (2423)	149 (6.15)	2274 (93.9)
5-14 years old (3350)	51 (1.52)	3299 (98.5)
>14 years old (7110)	54 (0.76)	7056 (99.2)

4.3 Associations between the Independent Variables and Diarrheal Disease

The Directly-Linked Dataset

The directly-linked dataset was created by merging the water quality and diarrheal disease datasets by linking the diarrheal disease data with the water quality data from the same observation week. Thus, any occurrences of diarrheal disease were assumed to be associated with the water quality during the same week (Figure 2). After the merge, 103 observations did not merge because they did not fit into the criteria for the merge variable. Thus, they were not included in the analysis.

T-tests were performed on all three independent variables, turbidity, total coliforms, and *E. coli*, to see if there was any difference between the means of groups with or without diarrhea.

Turbidity was the only independent variable that had a statistically significant difference between the two groups ($p = 0.005$) (Table 4).

All three variables were log-transformed to obtain a more normal distribution. Then, t-tests were performed on the log-transformed variables to see if the significance of any of the variables changed. However, turbidity was still the only variable that resulted in a significant difference between groups with diarrhea (Mean = 0.191) versus groups without diarrhea (Mean = 0.098; $p = 0.01$). Neither total coliforms nor *E. coli* levels were found to be significantly different between the two groups ($p = 0.716$; $p = 0.538$, respectively) (Table 4).

Table 4: T-test of Water Quality Variables and Diarrhea Status in the Directly-Linked Dataset

Variable	Non-Log-Transformed Values		Log ₁₀ -Transformed Values	
	Diarrhea	No Diarrhea	Diarrhea	No Diarrhea
Turbidity ^a				
<i>Mean</i>	3.10	2.19	0.191	.098
<i>p-value</i>	0.005		0.01	
Total Coliforms ^b				
<i>Mean</i>	1260	1272	2.74	2.72
<i>p-value</i>	0.894		0.716	
E. coli ^b				
<i>Mean</i>	178	153	1.30	1.25
<i>p-value</i>	0.484		0.538	

^a (NTU)

^b (MPN/100 mL)

^c (log₁₀NTU)

^d (log₁₀MPN/100 mL)

Binary logistic regressions were performed for all three log-transformed variables. Turbidity was the only variable that was significant for an association with diarrheal disease with an OR of 1.60 ($p = 0.010$) (Table 5). The odds of the occurrence of diarrheal disease were increased by a factor of 1.60 with a one unit increase in log turbidity. A one unit increase in log turbidity is the equivalent of a tenfold increase. Similar to the analyses of means, total coliforms and *E. coli* were found to not be associated with the occurrence of diarrheal disease.

All three log-transformed variables were then adjusted for age. After adjusting for age, none of the variables were found to be significant for an association with the occurrence of diarrheal disease (Table 5).

Table 5: Binary Logistic Regression of Individual Water Quality Variables and Diarrhea in the Directly-Linked Dataset

Variable	Unadjusted Values				Age-Adjusted Values			
	N	OR	p-value	CI (95%)	N	OR	p-value	CI (95%)
Turbidity ^a	6509	1.60	0.010	1.12- 2.30	6509	1.43	0.053	.995- 2.05
<i>E. coli</i> ^b	6514	1.05	0.538	0.893- 1.24	6514	1.14	0.130	.963- 1.34
Total coliforms ^b	6514	1.04	0.716	0.853- 1.26	6514	1.03	0.774	.846- 1.25

^a (\log_{10} NTU)

^b (\log_{10} MPN/100 mL)

The One-Week Lag Dataset

The one-week lag dataset was created by merging the water quality and diarrheal disease datasets by linking the water quality data with the diarrheal disease measurements taken one week after the collection of the water sample. Thus, this merge took into account the potential of an incubation time between exposure to the water sample and the occurrence of diarrheal disease (Figure 2). After the merge, 103 observations did not merge due to not meeting the determined merge criteria, so they were not included in the analysis.

T-tests were performed on all three independent variables, turbidity, total coliforms, and *E. coli*, to see if there was any difference between the means of groups with or without diarrhea. Turbidity was the only independent variable that had a statistically significant difference between the two groups ($p = 0.042$) (Table 6).

All three variables were log-transformed to obtain a more normal distribution. Then, t-tests were performed on the log-transformed variables to see if the significance of any of the variables changed. However, turbidity was still the only variable that resulted in a significant difference between groups with diarrhea (Mean = 0.205) versus groups without diarrhea (Mean = 0.094; $p = 0.005$). Neither total coliforms nor *E. coli* levels were found to be significantly different between the two groups ($p = 0.295$; $p = 0.624$, respectively) (Table 6).

Table 6: T-test of Water Quality Variables and Diarrhea Status in the One-Week Lag Dataset

Variable	Non-Log-Transformed Values		Log ₁₀ -Transformed Values	
	Diarrhea	No Diarrhea	Diarrhea	No Diarrhea
Turbidity^a				
<i>Mean</i>	2.87	2.16	0.205	0.094
<i>p-value</i>	0.042		0.005	
Total Coliforms^b				
<i>Mean</i>	1216	1271	2.63	2.71
<i>p-value</i>	0.570		0.295	
<i>E. coli</i>^b				
<i>Mean</i>	151	154	1.20	1.24
<i>p-value</i>	0.951		0.624	

^a (NTU)

^b (MPN/100 mL)

^c (log₁₀NTU)

^d (log₁₀MPN/100 mL)

Binary logistic regressions were performed for all three log-transformed variables. Turbidity was the only variable that was significant with an OR of 1.77 ($p = 0.005$; 1.19-2.62) (Table 7). As log turbidity increases by one unit, the odds of occurrence of diarrheal disease increases by a factor of 1.77. This suggests a positive association between diarrheal disease and turbidity. Similarly to the results of the T-tests, there was no association between the occurrence of diarrheal disease and either total coliforms and *E. coli*.

The three water quality variables were then adjusted for age. After adjustment, turbidity alone remained significant with an OR of 1.56. However, based on turbidity's p -value of 0.026, there was a slight decrease in significance compared to the unadjusted values (Table 7). The lack of significance for both total coliforms and *E. coli* remained unchanged by the age-adjustment.

Table 7: Binary Logistic Regression of Individual Water Quality Variables and Diarrhea in the One-Week Lag Dataset

Variable	Unadjusted Values				Age-Adjusted Values			
	N	OR	p-value	CI (95%)	N	OR	p-value	CI (95%)
Turbidity ^a	6243	1.77	0.005	1.19- 2.62	6243	1.56	0.026	1.05- 2.32
<i>E. coli</i> ^b	6248	0.956	0.624	0.798- 1.15	6248	1.04	.669	.868- 1.25
Total coliforms ^b	6248	0.904	0.296	0.748- 1.09	6248	.899	0.277	.742- 1.09

^a (\log_{10} NTU)

^b (\log_{10} MPN/100 mL)

The Monthly Average Dataset

The monthly average dataset was created by averaging all of the water quality measurements for each household for each month. These averaged water quality measurements were then linked with the corresponding diarrheal disease measurements for that month (Figure 3).

T-tests were performed on all three variables, turbidity, total coliforms, and *E. coli*, to see if there was any difference between the means of groups with or without diarrhea. Turbidity was the only variable that had a statistically significant difference between the two groups ($p < 0.001$) (Table 8).

All three variables were log-transformed to obtain a more normal distribution. Then, t-tests were performed on the log-transformed variables to see if the significance of any of the variables changed. However, turbidity was still the only variable that resulted in a significant difference between groups with diarrhea (Mean = 0.188) versus groups without diarrhea (Mean = 0.0946; $p < 0.001$). Neither total coliforms nor *E. coli* levels were found to be significantly different between the two groups ($p = 0.1405$; $p = 0.387$, respectively) (Table 8).

Table 8: T-test of Water Quality Variables and Diarrhea Status in the Monthly Average Dataset

Variable	Non-Log-Transformed Values		Log ₁₀ -Transformed Values	
	Diarrhea	No Diarrhea	Diarrhea	No Diarrhea
Turbidity^a				
<i>Mean</i>	2.99	2.19	0.188	0.0946
<i>p-value</i>	<0.001		< 0.001	
Total coliforms^b				
<i>Mean</i>	1332.98	1402.32	2.71	2.76
<i>p-value</i>	0.139		0.1405	
E. coli^b				
<i>Mean</i>	185.05	168.16	1.26	1.23
<i>p-value</i>	0.376		0.387	

^a (NTU)

^b (MPN/100 mL)

^c (log₁₀NTU)

^d (log₁₀MPN/100 mL)

Binary logistic regressions were performed between each of the three independent variables and diarrhea. Turbidity was the only variable that was significant with an OR of 1.73 ($p < 0.001$) (Table 9). Thus, the odds of the occurrence of diarrheal disease were increased by 1.73 as log turbidity increased by one unit. Neither total coliforms nor *E. coli* were found to be significant.

Adjustment for age was performed, which resulted in only turbidity being significant. Turbidity remained significant with a p-value of 0.001, and based on the OR, the odds of the occurrence of diarrheal disease increased by 1.45 as log turbidity increased by one unit. This resulted in a positive association between log turbidity and the occurrence of diarrheal disease.

Once again, neither total coliforms nor *E. coli* were found to be significant in this analysis (Table 9).

Table 9: Binary Logistic Regression of Individual Water Quality Variables and Diarrhea in the Monthly Average Dataset

Variable	Unadjusted Values				Age-Adjusted Values			
	N	OR	p-value	CI (95%)	N	OR	p-value	CI (95%)
Turbidity ^a	14912	1.73	< 0.001	1.38- 2.16	14912	1.45	0.001	1.15- 1.81
<i>E. coli</i> ^b	14931	1.05	0.387	0.945- 1.16	14931	1.10	0.067	.994- 1.21
Total coliforms ^b	14931	0.922	0.141	0.828- 1.03	14931	.901	0.065	.807- 1.01

^a (log₁₀NTU)

^b (log₁₀MPN/100 mL)

A multivariate logistic regression with all three water quality variables was conducted, adjusting for age. After running the analysis, both turbidity and total coliforms were found to be significant. For turbidity, the odds of diarrheal disease occurrence were 1.36 for every one unit increase of log turbidity (p-value = 0.012). This resulted in a positive association between turbidity and the occurrence of diarrheal disease (Table 10).

In contrast, there was a negative association between total coliforms and the occurrence of diarrheal disease. As log total coliforms increases by one unit, the odds of the occurrence of

diarrheal disease increased by 0.842 (p-value = 0.006). There was still not any significant association between the occurrence of diarrheal disease and *E. coli* levels (Table 10).

Table 10: Age-Adjusted Multivariate Logistic Regression of Water Quality Variables and Diarrhea in the Monthly Average Dataset (N = 14912)

Variable	OR	p-value	CI (95%)
Turbidity ^a	1.36	0.012	1.07-1.74
<i>E. coli</i> ^b	1.11	0.088	.984-1.25
Total coliforms ^b	0.842	0.006	.745-.953

^a (log₁₀NTU)

^b (log₁₀MPN/100 mL)

4.4 Summary of the Results

In this study, the water quality measurements and the diarrheal disease measurements were linked together in three different ways. For all of the t-test analyses, turbidity was the only water quality variable that showed a significant difference between groups with diarrhea and groups without diarrhea. Turbidity was found to have the greatest significance when the t-test was conducted on the monthly averages dataset (p-value < 0.001). However, this is most likely due to the larger sample size found in the monthly averages dataset (Table 11).

In the one-week lag dataset and the monthly averages dataset, age-adjusted univariate logistic regression analyses found a significant association between the turbidity levels and

diarrheal disease occurrence (Table 11). However, age-adjusted univariate logistic regression analysis performed on the directly-linked dataset did not find any significant associations between the occurrence of diarrheal disease and any of the water quality variables (Table 11). This lack of associations may be because this method of merging did not take the potential incubation time of diarrheal disease into account.

In contrast, the one-week lag dataset took potential incubation time into account by lagging the diarrheal disease dataset by one week. Thus, the one-week lag dataset did find a significant association between turbidity and diarrheal disease occurrence using age-adjusted univariate analysis. An OR of 1.56 was found for turbidity (p-value = 0.026) (Table 11).

Both the directly-linked and the one-week lag datasets only dealt with roughly half of the observations, which negatively impacts significance. In contrast, the monthly average dataset was comprised of all of the observations. When age-adjusted univariate logistic regression analysis was performed on this dataset, turbidity was still the only significant association with an OR of 1.45 (p-value = 0.001 (Table 11).

Due to the increased sample size and increased significance of the results of the univariate logistic regression analyses, a multivariate logistic regression analysis was only performed on the monthly averages dataset. The analysis was adjusted for age. After conducting this analysis, significant associations for both turbidity and total coliforms with the occurrence of diarrheal disease were found. For turbidity, a positive association was found (OR = 1.36; p-value = 0.012). However, total coliforms had a negative association with diarrheal disease occurrence (OR = 0.842; p-value = 0.006) (Table 11).

Based on these results, in this study population, turbidity was the only factor that potentially negatively impacted the health outcome by increasing diarrheal disease. For total

coliforms, there is evidence of a protective effect when controlling for turbidity, *E. coli*, and age (Table 11).

Table 11: Compilation of Results from T-tests, Age-Adjusted Univariate Logistic Regression, & Age-Adjusted Multivariate Regression of Individual Independent Variables versus Diarrhea

Variable	T-Test; Mean		Univariate Logistic Regression			Multivariate Logistic Regression*		
	Diarrhea (p-value)	No Diarrhea	OR	CI (95%)	p-value	OR	CI (95%)	p-value
Directly-Linked								
<i>Turbidity</i> ^a	0.191 (0.01)	0.098	1.43	.995-2.05	.053			
<i>Total Coliforms</i> ^b	2.74 (0.716)	2.72	1.03	.846-1.25	.774			
<i>E. coli</i> ^b	1.30 (0.538)	1.25	1.14	.963-1.34	.130			
One-Week Lag								
<i>Turbidity</i> ^a	0.205 (0.005)	0.094	1.56	1.05-2.32	.026			
<i>Total Coliforms</i> ^b	2.63 (0.295)	2.71	.899	.742-1.09	.277			
<i>E. coli</i> ^b	1.20 (0.624)	1.24	1.04	.868-1.25	.669			
Monthly Averages								
<i>Turbidity</i> ^a	0.188 (<0.001)	0.0946	1.45	1.15-1.81	.001	1.36	1.07-1.74	.012
<i>Total Coliforms</i> ^b	2.71 (0.1405)	2.76	.901	.807-1.01	.065	.842	.745-.953	.006
<i>E.coli</i> ^b	1.26 (0.387)	1.23	1.10	.994-1.21	.067	1.11	.984-1.25	.088

^a (log₁₀NTU)

^b (log₁₀MPN/10 mL)

* Multivariate logistic regressions were not performed on the directly-linked and the one-week lag datasets

Chapter V

DISCUSSION AND CONCLUSION

5.1 Discussion

The issue of water quality and its potential role in the occurrence of diarrheal disease has been a major public health issue since the beginning of the 20th century. While this issue has already been addressed in many developed countries through the creation of sophisticated water treatment measures, developing countries are still working towards reducing the impact of water quality on the health of their populations by improving access to clean water (Rheingans *et al.*, 2006). According to the most recent estimates, 884 million people still lack access to clean water, and 2.6 billion do not have adequate sanitation (JMP, 2010).

However, while it is accepted that water quality contributes to diarrheal disease, there is still not a lot of evidence available to corroborate this association between the levels of indicator pathogens and water quality. Previous studies have found different results depending on the population or the setting being studied (Moe *et al.*, 1991; Gundry *et al.*, 2004; Jensen *et al.*, 2004; Brown *et al.*, 2008). Thus, this study aimed to strengthen the evidence base for associations between water quality and diarrheal disease in a developing country.

Water Quality Dataset

Out of the study population, 31% of drinking water samples were reported to be treated prior to consumption. The most common treatment used was boiling. Thus, this means that the majority of the population is still drinking untreated water (68.8%) (Table 1).

Additionally, the total coliform and *E. coli* levels did not meet the WHO's water quality standards. WHO states that there should be no detectable *E. coli* or total coliforms in any 100 mL water sample (Havelaar *et al.*, 2001). After taking the inverse log, the majority of the *E. coli* levels were around 1 MPN/100 mL of water (Figure 7). While the *E. coli* levels are only slightly above the acceptable value, the total coliform levels vastly exceeded acceptable levels. The majority of the total coliform observations were clustered around 200 MPN/ 100 mL of water (Figure 6), which is not acceptable according to the WHO guidelines (Havelaar *et al.*, 2001).

However, the majority of the turbidity levels were below the acceptable level. According to the WHO, turbidity levels should be below 5 NTU (WHO, 2010). In this study, the levels ranged from 0.1 to 10 NTU with the majority of the observations being around 1 NTU after the inverse log was taken (Figure 5). Thus, it is surprising that turbidity was found to be significantly associated with diarrheal disease in the majority of analyses conducted on the merged datasets.

Additionally, during the 5-month study period, the turbidity levels and *E. coli* levels remained fairly constant throughout the entire time period with only very slight increases and decreases. In contrast, the total coliform levels peaked in November and December. All three indicators began to decline in January (Figure 8; Figure 9). It is not known if this is indicative of seasonality in the levels of the indicators because the study was not long enough to be able to determine this.

However, it is possible that there is some seasonality among the different indicators. In the study conducted by Ako Ako *et al.*, they noticed a seasonal pattern to the occurrence of diarrheal disease in Cameroon due to rainfall (Ako Ako *et al.*, 2009). Although the Ako Ako *et al.* (2009) study looked at diarrheal disease occurrence, it is possible that the amount of rainfall could influence the levels of turbidity, total coliforms, and *E. coli*, which could then influence

the levels of diarrheal disease. Further studies with longer time periods are needed to determine if seasonality is a factor in the levels of indicators in the Dominican Republic.

Diarrheal Disease Dataset

During the course of the five-month study period, there were 430 reported incidences of diarrheal disease (3.02%) (Table 2). After adding together the categories of under 2 years old and 2-4 years old, it was shown that the majority of the cases (325) occurred in children under 4 years old. As the population increased in age, the frequency of diarrhea among the population decreased (Table 3).

The results of this study are similar to what has been shown before in that children under five years old are disproportionately affected by diarrheal disease. UNICEF estimates that around 2.5 billion cases of diarrhea occur annually among children under five years old, which results in 1.5 million deaths occurring among this population. Additionally, the incidence of diarrheal disease is the highest among children 2 years old and younger (UNICEF, 2009). Unfortunately, the morbidity from diarrheal disease among young children has remained relatively constant since the 1980s despite large-scale efforts to improve health (O’Ryan *et al.*, 2005; UNICEF, 2009).

Associations Between the Water Quality Variables and Diarrheal Disease

For two out of the three data configurations analyzed in this study, turbidity was the only water quality variable that was found to be statistically significant with a positive association

between increasing turbidity levels and the increasing occurrence of diarrheal disease. This was somewhat surprising given that the majority of the turbidity levels of the water samples (Figure 5) were below the WHO's standards (Havelaar *et al.*, 2001). However, these results do support previous findings. In the systematic literature review conducted by Mann *et al.* (2007), the occurrence of an association between turbidity and gastrointestinal disease was dependent on the population and setting being studied (Mann *et al.*, 2007). Additionally, Tinker *et al.* (2010) conducted a study in Atlanta that found a slight association between an increased number of emergency room visits for gastrointestinal illness and raw water turbidity (Tinker *et al.*, 2010).

Both of these studies (Mann *et al.*, 2007; Tinker *et al.*, 2010) found only slight associations or inconsistent associations among developed countries. However, these two studies dealt with populations in developed countries with improved water sources. Thus, while these studies may have found only small associations or inconsistent associations in developed countries, it is logical to assume that in a developing country with decreased access to improved water supplies, greater associations may be found. This supposition was supported by the finding of a significant association between turbidity and occurrence of diarrheal disease in this study. This association remained even after controlling for *E. coli*, total coliforms and age (Table 14).

Total coliforms were only found to be significant after controlling for turbidity, *E. coli*, and, age. Interestingly, this study found that total coliforms had a protective effect. Thus, according to these results, increased levels of total coliforms result in decreased occurrence of diarrheal disease. However, there are not a lot of previous findings that support a protective effect for total coliforms. In 1987, a study conducted by Zmirou *et al.* found that fecal coliforms were protective in small villages of fewer than 400 people. In communities with more than 400 people, they found that fecal coliforms did not affect the occurrence of gastrointestinal disease

(Zmirou *et al.* 1987). However, this study included communities in the Dominican Republic that were larger than 400 people, so the findings are not supported by previous literature.

It is also possible that the total coliform levels were limited by the detection limits of the assay as well as the non-Gaussian distribution of the samples. Thus, this limitation could result in the appearance of a protective effect when no such effect exists. In order to confirm or refute the existence of a protective effect, future studies need to use categorical total coliform levels in the analyses.

E. coli was not significant in any of the merges. This finding does support previous findings that showed that the levels of *E. coli* in drinking water are not indicative of the occurrence of diarrheal disease (Jensen *et al.*, 2004; Gundry *et al.*, 2004). However, a large number of the *E. coli* observations were clustered around 0 log₁₀MPN/100 mL, which is around 1 MPN/100 mL (Figure 7). In the studies conducted by Moe *et al.* (1991) and Brown *et al.* (2008), *E. coli* levels were only significant at higher concentrations of *E. coli* (Moe *et al.*, 1991; Brown *et al.*, 2008).

5.2 Study Limitations

One of the potential limitations to this study is the use of two different datasets to create one comprehensive dataset. Ultimately, three different data merges were conducted because of the differences in the timeframes of each study. The first two merges merged the biweekly water quality dataset with the weekly diarrheal disease dataset. Because of the nature of the merges, both of the resultant merged datasets contained only half of the observations. Although the loss of half of the data was expected, it still could have a negative impact on the level of significance of any of the independent variables.

Additionally, during the first two merges, some data from the water quality dataset did not merge with the diarrheal disease dataset for unknown reasons. This resulted in the loss of 103 observations due to the merge criteria used. Future studies would need to create more comprehensive merge criteria.

The final merge combined the biweekly water quality measures with the weekly diarrheal disease dataset. In order to include all of the observations, all of the water quality measurements from each month were averaged together resulting in one monthly water quality measurement. This monthly averaged value was then assigned to all of the household observations from that month.

Although averaging the water quality data was necessary to be able to combine both datasets in their entirety, it also creates another potential limitation. By averaging the values, the resultant value can potentially be skewed by outliers. Thus, it is difficult to ascertain if there were any extremely high or low levels of any of the variables that might correspond with incidences of diarrheal disease. Additionally, the differences between a water source that was only used by children versus one that was only used for adults cannot be measured. It is possible that the differences between the types of treatment used for the different sources may impact the levels of indicators.

Another possible source of limitations was in the collection of the data. All of the diarrheal disease data was dependent on one member of the household answering questions about the entire family's diarrheal disease history for the past week. Thus, the respondent's recall of diarrheal disease occurrence for their entire family may not be completely accurate. Additionally, the use of family units created clustering in the study. Due to repeated observations within the same households and due to the potential for family members to inadvertently affect

the responses of other family members, clustering was a factor in this study. Future studies need to take clustering into account during the conducting of the statistical analyses.

Finally, this study looked at the odds ratios for all of the indicator levels. Thus, the data is only able to suggest an association or lack of an association. It is not possible to establish causality. Therefore, it is always possible that the diarrheal disease was caused by factors outside of the scope of this study.

5.3 Recommendations

Future studies of these datasets need to be conducted to look at some of the unanswered questions and to further clarify the results presented in this study. The *E. coli* and total coliform levels need to be stratified into different categories and then look at the possible associations between them and diarrheal disease. It has been seen in two different studies (Moe *et al.*, 1991; Brown *et al.*, 2008) that *E. coli* only becomes significantly associated with diarrheal disease occurrence at higher levels of *E. coli*. Thus, this possibility needs to be done in future studies.

The association between turbidity levels and diarrheal disease occurrence needs to be further explored. Finally, the possibility of total coliforms causing a protective effect needs to be further studied using categorized total coliform levels.

5.4 Conclusion

Overall, this study further strengthened the evidence that young children suffer the most from diarrheal disease in developing countries. It also showed that even though 95% of the

population of the Dominican Republic have access to improved water sources (WHO, 2010), the levels of total coliforms and *E. coli* in some areas' drinking water are not considered potable by WHO's standards (Havelaar *et al.*, 2001). Additionally, the results of this study showed that among this population the turbidity levels are potentially a significant factor in the occurrence of diarrheal disease, despite the acceptable turbidity levels in the drinking water. Finally, this study showed that total coliform levels may confer a protective effect.

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